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Proc. R. Soc. B published online 7 November 2012
doi: 10.1098/rspb.2012.2359

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Cite this article: Krkošek M, Revie CW, Gargan PG, Skilbrei OT, Finstad B, Todd CD. 2012 Impact of parasites on salmon recruitment in the Northeast Atlantic Ocean. *Proc R Soc B* 20122359. <http://dx.doi.org/10.1098/rspb.2012.2359>

Received: 4 October 2012

Accepted: 15 October 2012

Subject Areas:

ecology, health and disease and epidemiology

Keywords:

recruitment, salmon, parasites, conservation, fisheries

Author for correspondence:

Martin Krkošek

e-mail: martin.krkošek@otago.ac.nz

Electronic supplementary material is available at <http://dx.doi.org/10.1098/rspb.2012.2359> or via <http://rspb.royalsocietypublishing.org>.

Impact of parasites on salmon recruitment in the Northeast Atlantic Ocean

Martin Krkošek¹, Crawford W. Revie², Patrick G. Gargan³, Ove T. Skilbrei⁴, Bengt Finstad⁵ and Christopher D. Todd⁶

¹Department of Zoology, University of Otago, Dunedin, New Zealand

²Atlantic Veterinary College, University of Prince Edward Island, Charlottetown, Canada

³Inland Fisheries Ireland, Dublin, Ireland

⁴Institute of Marine Research, Bergen, Norway

⁵Norwegian Institute for Nature Research, Trondheim, Norway

⁶Scottish Oceans Institute, University of St Andrews, St Andrews, UK

Parasites may have large effects on host population dynamics, marine fisheries and conservation, but a clear elucidation of their impact is limited by a lack of ecosystem-scale experimental data. We conducted a meta-analysis of replicated manipulative field experiments concerning the influence of parasitism by crustaceans on the marine survival of Atlantic salmon (*Salmo salar* L.). The data include 24 trials in which tagged smolts (totalling 283 347 fish; 1996–2008) were released as paired control and parasiticide-treated groups into 10 areas of Ireland and Norway. All experimental fish were infection-free when released into freshwater, and a proportion of each group was recovered as adult recruits returning to coastal waters 1 or more years later. Treatment had a significant positive effect on survival to recruitment, with an overall effect size (odds ratio) of 1.29 that corresponds to an estimated loss of 39 per cent (95% CI: 18–55%) of adult salmon recruitment. The parasitic crustaceans were probably acquired during early marine migration in areas that host large aquaculture populations of domesticated salmon, which elevate local abundances of ectoparasitic copepods—particularly *Lepeophtheirus salmonis*. These results provide experimental evidence from a large marine ecosystem that parasites can have large impacts on fish recruitment, fisheries and conservation.

1. Introduction

Infectious diseases may threaten biodiversity conservation and food security [1,2]. Conceptually, a number of disease-associated mechanisms may elevate extinction risk, which include small pre-epidemic population size and the presence of a reservoir host population [3]. Furthermore, evolutionary similarity to domesticated animals may be a key factor associated with parasite-mediated population declines of wildlife [4]. For food security, disease threats to domesticated populations are associated with increased host density or emergence of novel pathogen strains [1,5]. More broadly, parasites are increasingly being recognized for their direct and indirect effects on host population dynamics and community structure [6–10].

However, clear experimental demonstrations of parasites regulating or limiting host population dynamics in field conditions are rare [11,12]. For marine fishes, uncertainty may be particularly high because fish recruitment is highly stochastic and nonlinear [13,14], and analyses of recruitment are correlative with correlates often being ephemeral [15]. Nevertheless, there is a growing perception of the potential regulatory importance of parasites in marine fish populations [16–19]. Furthermore, parasites that are shared between wild marine fish populations and evolutionarily similar domesticated stocks in

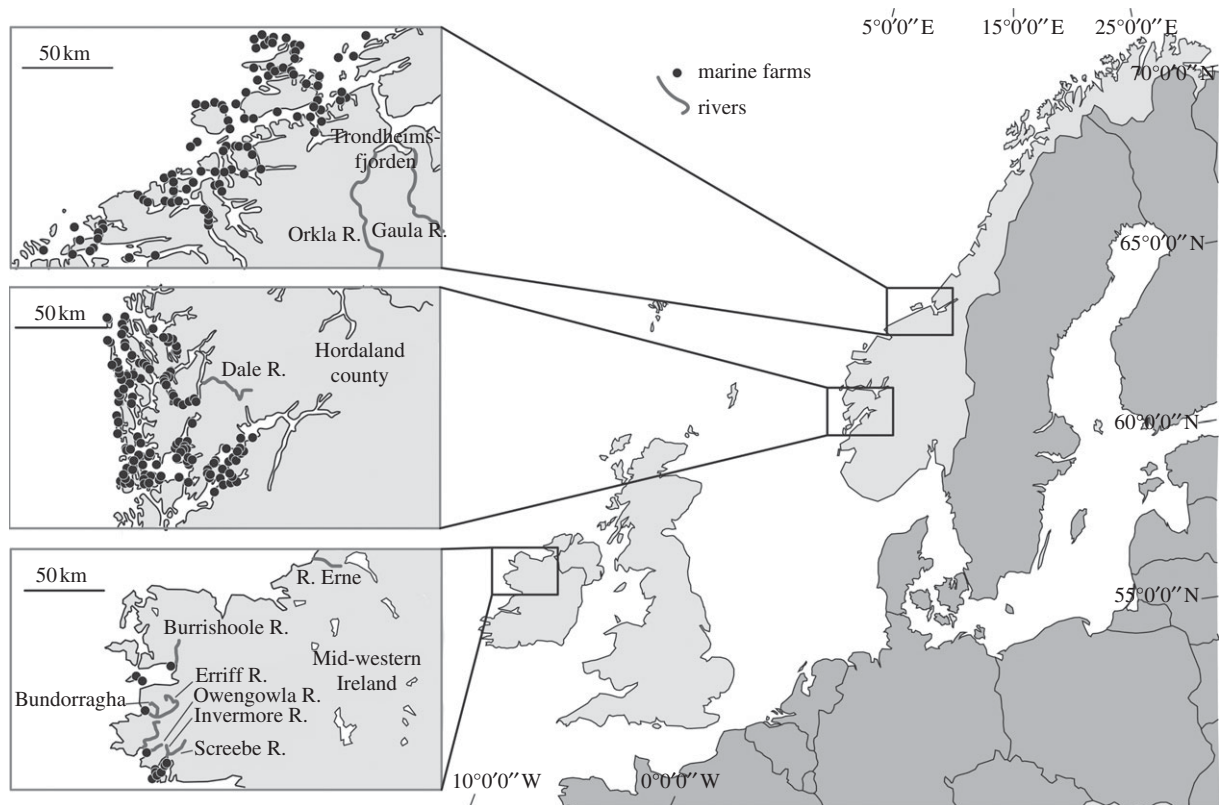


Figure 1. Map of the study regions showing rivers where salmon smolts were released and salmon farm locations (black circles). Interpretation of farming activities over the course of the study period should be made with caution, because the distributions of farms, stocking levels, parasite levels and management approaches have varied over the years.

aquaculture may affect fish population dynamics, fisheries productivity, biodiversity conservation and aquaculture productivity [5,20,21].

In coastal seas, domesticated populations in aquaculture are rapidly growing and these experience both incidental and persistent disease outbreaks [22–24]. Although disease and its control are costly for aquaculture producers, concerns also have emerged on the potential wider consequences for marine ecosystems and wild fisheries [20,21,25,26]. Perhaps nowhere is this more contentious than with Pacific (*Oncorhynchus* spp.) and Atlantic salmon (*Salmo salar* L.) and their native, specialist ectoparasitic copepod, the salmon louse *Lepeophtheirus salmonis* (Krøyer) [27–29]. In the North Atlantic, stocks of wild *S. salar* have shown a marked and steady decline of approximately 45 per cent since the 1980s, with variability among stocks [30], and correlations to salmon aquaculture have been drawn [31,32].

Lepeophtheirus salmonis is a marine ectoparasitic copepod of salmonids (*Salmo* spp., *Salvelinus* spp., *Oncorhynchus* spp.) that feeds on host surface tissues causing morbidity and mortality [27,28]. There is no intermediate host and the parasitic gravid female releases free-living larvae to the water column where they can be dispersed among wild and farmed populations [5,33]. *Lepeophtheirus salmonis* is extremely unusual amongst macroparasites [34,35] in typically reaching 100 per cent prevalence on wild one sea-winter (1SW) and two sea-winter (2SW) Atlantic salmon [36,37]. Persistent re-infestation of wild salmon at sea is apparent from the significantly higher mean abundance of *L. salmonis* on wild 2SW fish compared with 1SW adults [37]. Thus, *L. salmonis* must be viewed either as being exceptionally well adapted in locating the host fish or, more likely, the behaviour of salmon is such that they inevitably

encounter the infective copepodid stage of *L. salmonis* both in coastal waters and in the open Atlantic ocean.

Numerous studies have associated salmon farming and sea louse infestations of wild juvenile salmon, but the impact on salmon recruitment remains contentious [20,38–44]. A primary reason is that many studies are correlative and results depend on modelling assumptions [32,38,39]. By contrast, manipulative field experiments provide an opportunity for strong inferences based on differential survival between paired releases of control and parasiticide-treated groups of emigrant juvenile salmon smolts [40,42,45,46]. The parasiticides affect crustaceans, which include other native salmon parasites, particularly *Caligus elongatus*, though *L. salmonis* remains the focus owing to its prevalence and adverse lethal and sub-lethal effects [27,28]. Some of the field experiments have identified significant impacts on salmon survival [42,45], whereas others have claimed no significant effect at the population level [40,46].

We assembled and analysed all the published results of large-scale, pairwise field experiments of the marine survival of Atlantic salmon in the Northeast Atlantic Ocean. These studies were conducted in Ireland and Norway, and typically occurred in river systems discharging to near shore marine waters that accommodated salmon farms [40,42,45–47]. Figure 1 shows the location of coastal salmon farm sites adjacent to the named experimental release rivers in Norway and Ireland; however, not all farm sites were in active production in all years of experimental releases. The published studies all were substantial, involving the release of thousands of hatchery-raised juvenile salmon smolts into rivers. Prior to release, all fish were tagged (either coded wire or carlin tags) and approximately half the juveniles received a parasiticide treatment, while the others

Table 1. Summary of the data, giving the year, country and river of release as well as the numbers of smolts released (N) and of one sea-winter (1SW) adults recovered (R) for treatment groups receiving parasiticides (t) and controls (c). (Also shown are the total number of smolts released (N_{total}) and sources of the data (ref.), those studies with multiple within-season releases that were grouped (*), and the studies that used the proprietary Substance Ex as treatment ([†]) rather than emamectin benzoate. Note that numbers of recaptured salmon for references [40] and [46] are their survival estimates that incorporated raising factors applied to the raw data (tags recovered), and may therefore be inflated relative to the other data.)

year	country	river	N_t	R_t	N_c	R_c	N_{total}	ref.
1996	Norway	Agdenes	3000	53	2985	43	5985	[47] [†]
1997	Norway	Agdenes	2935	16	2936	12	5871	[47] [†]
1998	Norway	Agdenes	2966	37	2977	14	5943	[47] [†]
2002	Norway	Dale	5086	56	4859	40	9945	[45]*
2004	Ireland	Erriff	4325	44	4229	34	8554	[42]
2005	Ireland	Erriff	4659	37	4689	2	9348	[42]
2003	Ireland	Invermore	4589	17	4594	9	9183	[42]
2004	Ireland	Invermore	4653	37	4671	26	9324	[42]
2005	Ireland	Invermore	4716	31	4750	17	9466	[42]
2003	Ireland	Owengowla	4955	35	4822	3	9777	[42]
2004	Ireland	Owengowla	4655	51	4699	22	9354	[42]
2005	Ireland	Owengowla	4583	54	4735	53	9318	[42]
2001	Ireland	Burrishoole	5496	565	10039	992	15535	[40]
2002	Ireland	Burrishoole	5960	544	5989	545	11949	[40]
2003	Ireland	Burrishoole	4755	472	4587	374	9342	[40]
2004	Ireland	Burrishoole	4437	402	4369	398	8806	[40]
2005	Ireland	Burrishoole	3793	253	3867	182	7660	[40]
2006	Ireland	Burrishoole	8716	508	12779	549	21495	[40]*
2007	Ireland	Burrishoole	6746	492	6795	435	13541	[40]
2008	Ireland	Burrishoole	10132	163	10224	125	20356	[40]*
2001	Ireland	Bundorragha	12787	2056	12753	1873	25540	[46]*
2006	Ireland	Erne	5752	70	10357	68	16109	[46]
2006	Ireland	Lee	5207	10	5131	10	10338	[46]
2006	Ireland	Screebe	10990	157	9618	121	20608	[46]

were untreated controls. Targeted tag recovery programmes and screening of commercial catches of return adult spawners allowed comparisons of marine survival between treatment and control groups. The large majority of the recovered fish spent 1 year at sea, and we focus on these so-called 1SW fish in the main analysis. Further details, including comparative analyses of return adults of all sea ages (1SW, 2SW and 3SW), and the data analysed, are presented in the electronic supplementary material.

2. Material and methods

(a) Data sources

The data spanned 24 trials between 1996 and 2008, and included a total of 283 347 smolts released into 10 rivers mostly in Ireland but also Norway (figure 1 and table 1; electronic supplementary material). Most trials involved a single release of paired control–treatment groups into a river (or estuary [47]) during the spring emigration (typically April–May). In some trials, there were multiple separate releases within an emigration season: for these, we summed the data across the multiple releases into a single observation for that annual trial. In 21 of 24 trials, the in-feed treatment emamectin benzoate (Slice; Schering-Plough) was applied; this is

a widely used parasiticide in salmon aquaculture [48]. For three trials, a different proprietary compound, Substance Ex (Alpharma), was used [47] (table 1), which involves a topical bath treatment.

(b) Analytical approaches

We considered three independent but complementary analyses based on odds ratios, paired sample t -tests, and mixed-effects models, each of which synthesize the data across studies to quantitatively assess sea louse-induced mortality of Atlantic salmon. The first approach is a standard survival analysis that leads to estimates of the odds ratio in survival between control–treatment data pairs, as well as an overall meta-analytic mean odds ratio [49]. The second analysis begins with the standard assumptions of survival analysis, and leads to a simple paired sample t -test of survival estimates on a natural log scale. A key advantage of this approach is that it permits an estimation of the per cent of adult salmon recruitment that is lost to parasites. The third analytical approach also begins with the standard assumptions of survival analysis but, relative to the second approach, differs subtly in how residual variation is distributed between control and treatment groups. This third approach leads to a mixed-effects model of survival estimates, again on a log scale, and also permits estimation of the per cent of salmon recruitment lost to parasites.

Our initial standard meta-analysis is an approach that is often adopted when seeking to estimate the overall effect associated with a range of trials that report the same outcome variable. Because the outcomes were dichotomous, an appropriate measure of association was the odds ratio. Given the degree of heterogeneity evident among the studies, a random effects meta-analysis [49] was adopted. The meta.DSL function from the rmeta package of the R statistical software was used to carry out the analysis.

The second and third approaches begin with a standard survival analysis where the probability of survival from time $t = 0$ to time $t = T$ is

$$s = \exp\left(-\int_0^T \mu(t) dt\right), \quad (2.1)$$

and μ is known variously as the instantaneous mortality rate, hazard rate or force of mortality [50].

The second analytical approach involves a model of the data according to

$$\left. \begin{aligned} s_{C,i} &= \exp\left(-\int_0^T \mu_C(t) dt + \theta_i + \varepsilon_{C,i}\right) \\ s_{P,i} &= \exp\left(-\int_0^T \mu_P(t) dt + \theta_i + \varepsilon_{P,i}\right) \end{aligned} \right\}, \quad (2.2)$$

where $s_{C,i}$ and $s_{P,i}$ are the proportions of fish that were recovered from the control group (C) and parasiticide-treatment group (P) in trial i . The parameters μ_C and μ_P are the instantaneous mortality rates of Atlantic salmon from control (C) and parasiticide-treatment (P) groups. The mortality rate then is integrated over the duration of the study from the time of release ($t = 0$) to the average time of recapture ($t = T$). Variation in mortality rates among trials (but not between control and treatment groups within trials) is modelled as the normally distributed random variable, θ_i with a mean of zero; thus, the treatment and control groups in the same release share the same environmental variability in mortality rate. The parameter θ_i also encompasses random variation in T among trials, but we assume that T does not vary between control and treatment groups within trials. The parameter $\varepsilon_{C,i}$ is a normally distributed random variable with a mean of zero that models random variation among trials in parasite exposure in control groups. The parameter $\varepsilon_{P,i}$ is a normally distributed random variable with a mean of zero that models random variation among trials in the level of protection conferred by parasiticide treatment.

Equations (2.2) can be simplified to

$$\left. \begin{aligned} \ln(s_{C,i}) &= \beta_C + \theta_i + \varepsilon_{C,i} \\ \ln(s_{P,i}) &= \beta_P + \theta_i + \varepsilon_{P,i} \end{aligned} \right\}, \quad (2.3)$$

where

$$\left. \begin{aligned} \beta_C &= -\int_0^T \mu_C(t) dt \\ \beta_P &= -\int_0^T \mu_P(t) dt \end{aligned} \right\}. \quad (2.4)$$

Equations (2.4) lead to an estimate of the average mortality owing to parasites

$$M = 1 - \exp[-\varphi], \quad (2.5)$$

where $\varphi = \beta_P - \beta_C$. It follows that φ can therefore be estimated as

$$\ln(s_{i,P}) - \ln(s_{i,C}) = \varphi + \varepsilon_i, \quad (2.6)$$

where $\varepsilon_i = \varepsilon_{P,i} - \varepsilon_{C,i}$ is a random normal variable with mean of zero and variance to be estimated. Equation (2.6) indicates that statistical inference on φ , and therefore parasite-induced

mortality of Atlantic salmon, can be conducted as a simple paired sample t -test between paired survival estimates of control and treatment groups among trials on a natural log scale.

The third model of the data we consider is

$$s_{i,j} = \exp\left(-\int_0^T \mu_j(t) dt + \theta_i + \varepsilon_{i,j}\right), \quad (2.7)$$

where j represents control (C) or parasiticide-treatment (P) groups. Equation (2.7) is similar to equations (2.2), except that $\varepsilon_{i,j}$ is a normally distributed random variable that is distributed over control and treatment groups across the dataset. Here, it is assumed that mortality rates for each individual group—be it control or treatment—vary among each other according to a single random variable, whereas in equations (2.2) variation in mortality rate among trials within treatment or control groups was modelled separately for control and parasiticide-treatment groups.

This third approach leads to a model for the log-transformed survival data

$$\ln(s_{i,j}) = \beta_j + \theta_i + \varepsilon_{i,j}, \quad (2.8)$$

which has the form of a mixed-effects model, where θ_i is a random effect for shared environmental variability for control and treatment groups within trials and $\varepsilon_{i,j}$ is the residual variation. Here, statistical inference on whether parasiticide treatment has a significant effect on survival can be conducted by means of a likelihood ratio test between equation (2.8) and its null version where β_j is constant ($\beta_j = \beta_0$). Estimation of φ , and therefore parasite-induced mortality (M) as defined above, can be conducted from the maximum-likelihood estimate of the difference (and associated s.e.) between the treatment and control fixed effects in the mixed model (equation (2.8)).

3. Results

Meta-analysis of differential survival between control and parasiticide-treatment groups indicated a consistent directionality of the effect of treatment among trials, albeit with variation among trials in effect size, and a significant protective effect of treatment (overall odds ratio 1.29 with 95% CI: 1.18–1.42; figure 2). The paired sample t -test revealed significant differences in survival between treatment and control groups ($t_{23} = 3.37$, p -value = 0.003). Similarly, for our third (mixed model) analytical approach, treatment had a significant effect on survival (likelihood ratio test; $\chi^2_1 = 9.64$, $p = 0.002$). The estimated difference in log survival between paired treatment and control groups was $\varphi = 0.50$ from both the latter analyses. Uncertainty in φ was given by 95% CIs of 0.19–0.80 in the paired sample t -test, and a range of two standard errors giving the interval 0.20–0.79 in the mixed model. These estimates of φ correspond to parasite-associated mortality of 1SW Atlantic salmon of $M = 1 - \exp[-\varphi] = 39.3$ per cent (95% CI from paired sample t -test, 17.5–55.3%; range of two standard errors from mixed model, 18.3–54.8%). The complementary analysis (see the electronic supplementary material) that included all sea ages (1–3SW) of returning adults gave a final estimate of $M = 39.9$ per cent.

4. Discussion

These results provide manipulative experimental evidence at a large spatial scale from the marine environment that parasitism is a significant limiting factor for the recruitment of Atlantic salmon. Treatment of smolts with parasiticide

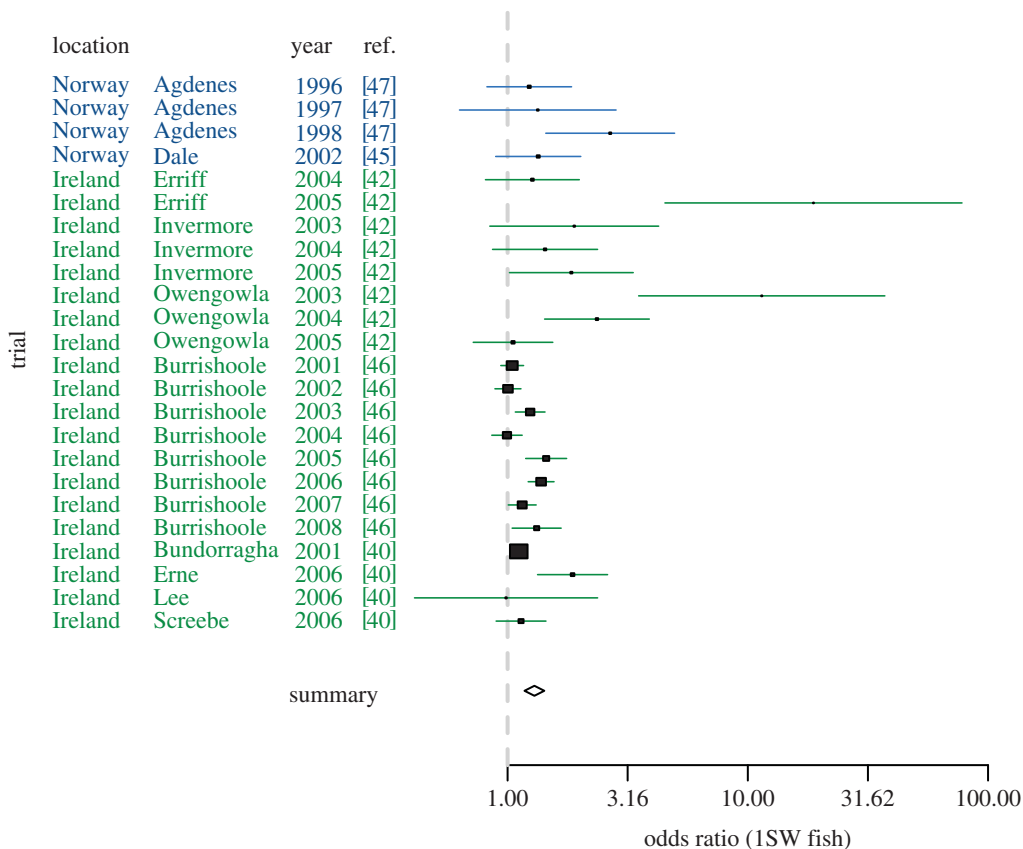


Figure 2. Forest plot displaying a random effect meta-analysis of the effect of treatment on the likelihood of a one sea-winter (1SW) adult salmon returning. Horizontal lines represent the 95% CIs of the effect size in each trial, and the relative sizes of solid squares reflect the percentage weighting (based on s.e. of effect sizes) of each trial in the meta-analysis (range 0.42–8.12%). The open diamond shows the overall meta-analytic effect across all studies, with its width corresponding to the 95% CI. Results are given by trial, identified by location (country and river of release of smolts), year (year when smolts were released) and ref. (the reference where the data were published).

significantly improved their survival to adulthood relative to control untreated groups. The most common treatment in the trials, emamectin benzoate, has a half-elimination period of 9–12 days in Atlantic salmon smolts [51] and is effective for up to the first 1 to 2 months at sea [48]. The differential survival between control and treatment groups is therefore attributable to near shore and coastal waters where the juvenile post-smolts transit from rivers to offshore waters. These near shore waters typically host large domesticated salmon populations in aquaculture operations (figure 1), which represent a large—but not exclusive—source of sea lice. Potential wild sources of sea lice include sea trout (*Salmo trutta* L.), Arctic char (*Salvelinus alpinus* L.) and spring-returning adult *S. salar* [27].

Although we focus attention on *L. salmonis*, in view of its prevalence and known adverse effects on host salmon, it is likely that *C. elongatus*, which also is native and occurs on both wild and farmed salmon [27], contributes to mortality. In addition, it is not implausible that there may be other unknown effects of the treatments that improve survival. Another caveat is that the data are based on hatchery-reared smolts, which, among other differences with wild smolts, are larger: hatchery smolts are typically 18–19 cm, whereas wild smolts are typically 11–14 cm in Ireland and Norway. The larger size of hatchery smolts can partially offset their typically low marine survivorship compared with wild smolts [52]. Nonetheless, owing to the size-related effects of sea-louse-induced stress on smolts [27,28], wild

salmon would probably experience higher mortality than we have estimated here for hatchery smolts.

We considered the possibility of a dose-dependent effect—i.e. including sea louse abundance on farms as a covariate—but proceeded otherwise. It was not possible to formally structure the analysis in this way because we lack sufficient farm data for all areas and years. Furthermore, local hydrodynamic processes that disperse the infective sea louse larval stage will be highly variable among locations [33,53], making it difficult to construct a covariate that is geographically comparable. Rather, we regard the variance in the data of differential survival as representative of the spatio-temporal variation of ectoparasite exposure that wild salmon experience in salmon-farming regions of the Northeast Atlantic.

While it is apparent that parasites have potential to be a significant source of mortality in wild fish populations [10,16,54], it is difficult to measure marine mortality associated with disease, and furthermore to evaluate whether such mortality scales up to a limiting or regulating factor of recruitment. Disease-associated mortality of fishes in the marine environment often goes undetected or is observed sporadically, and parasite-induced mortality may be compensatory, for example owing to predation [55]. Some examples include parasite-associated mortality of juvenile coho salmon (*Oncorhynchus kisutch*) with infestations of the digenean trematode *Apophallus* sp. [56], mortality of European eels (*Anguilla anguilla*) infested with the nematode *Anguillicoloides* (= *Anguillicola*) *crassus* [57], and mortality of

juvenile Pacific salmon (*Oncorhynchus* spp.) owing to the trematode *Nonophyetus salmincola* [58].

However, none of the above example studies of parasite-related mortality have been able to establish if parasitism is linked with declines in host fish recruitment, or is otherwise offset by compensatory mechanisms. From meta-analysis of taxonomically diverse field- and laboratory-based experimental studies, parasitized hosts tend to be approximately 2.6 times more likely to die compared with control groups, and that odds ratio is likely to be higher for fishes [10]. It has to be acknowledged, however, that variation in the odds ratio of host mortality is influenced by both taxon and latitude, and differs for (direct) parasite-associated effects and (indirect) predation-mediated influences of parasites on host survivorship [10]. Our analysis for one species of fish yielded a meta-analytic mean odds ratio of 1.29 (95% CI: 1.18–1.42), which is relatively low compared with reported values [10], and yet which still implies a high proportion (39%) of host mortality attributable to a specialist ectoparasite.

In contrast to the foregoing [56–58], our analysis of experimental manipulations of emigrant Atlantic salmon smolts permitted an explicit, though indirect, quantitative test of the significance of host mortality attributable to a specialist ectoparasite. All experimental fish were known to be free of *L. salmonis* infection at the start of the experiment, because the parasite is acquired only following seawater entry by the emigrant smolt. Fish in the control and treated groups all were identifiable by tags, and surviving adults that successfully completed the return marine migration in subsequent years could be recaptured because of the fidelity of salmon to their natal (or experimental release) river. Moreover, both the prophylactic experimental treatments applied have specific effects in controlling arthropod ectoparasites and emamectin benzoate is widely used in the aquaculture industry specifically to control *L. salmonis* infestations on cultured salmon.

The three analytical approaches we used reveal a key insight into interpreting the results of salmon and sea lice studies that formed the basis of our meta-analysis. While the assembled data indicate a clear consistency in direction

of parasiticide effects on survival, the magnitude of the odds ratio is relatively small. This is owing to the very high marine mortality that naturally affects both control and treatment groups. This, perhaps, underlies the conclusions of some studies that the effect is biologically insignificant relative to the overall mortality within a salmon life cycle [40,46]. However, our analyses based on paired sample *t*-tests and mixed-effects models allowed us to account for the high natural mortality and isolate the estimated loss of recruitment owing to parasitism, revealing a large effect of parasites. Precisely because natural mortality rates are high, even a proportionally small additive mortality from parasites can amount to a large loss in salmon recruitment.

It is rare to identify and quantify the factors that affect population dynamics of marine fishes [15] as well as the role of infectious disease in conservation [59]. Our results indicate that parasite-associated mortality may cause the closure of some fisheries when conservation targets of return adult abundances are not being met. However, the implications of our results may be most acute for small populations in small river systems. Due largely to the fidelity to their natal rivers, populations of Atlantic salmon typically show substantial genetic structuring and variability that is considered adaptive [60]. Small river systems that support salmon populations of low effective population size [61] will be especially vulnerable. The concern therefore is not only for a 39 per cent loss in salmon abundance, but also the loss of genetic variability and its associated potential for adaptation to other environmental changes. More broadly, and in contrast to the conclusions from two of the original studies [40,46], our results supply manipulative field experimental evidence at a large spatial scale that parasitism may be a significant limiting factor for marine fishes, fisheries and conservation.

We are grateful to Brendan Connors for comments on the manuscript. This work was supported by funding from the University of Otago, the Norwegian Directorate for Nature Management, the Canada Research Chair Programme at UPEL, the Norwegian Institute of Marine Research and EU Framework 5 (Q5RS-2002-00730).

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